

Ecology

The Macroalgae Ecological Quality Index (MEQI) in the Basin of Mexico: a proposal of aquatic bioindicators for peri-urban rivers

El Índice de Calidad Ecológica de Macroalgas (MEQI) en la cuenca de México: una propuesta de bioindicadores acuáticos en ríos periurbanos

Javier Carmoma-Jiménez ^a, Victor Hugo Salinas-Camarillo ^{a, b}, Angela Caro-Borrero ^{a, *}

^a Universidad Nacional Autónoma de México, Facultad de Ciencias, Departamento de Ecología y Recursos Naturales, Laboratorio de Ecosistemas de Ribera, Circuito Exterior s/n, Ciudad Universitaria, Coyoacán, 04510 Mexico City, Mexico

^b Universidad Nacional Autónoma de México, Posgrado en Ciencias del Mar y Limnología, Av. Ciudad Universitaria 3000, Coyoacán, 04510 Mexico City, Mexico

*Corresponding author: angelacaro@ciencias.unam.mx (Angela Caro-Borrero)

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Abstract

We assessed macroalgal responses to anthropogenic pressures in the Basin of Mexico to examine whether these biological indicators exhibited a discriminatory and explanatory correlation with ecological quality and propose the “Macroalgae Ecological Quality Index” (MEQI). The proposal included 3 steps: *i*) characterization of trophic level and hydromorphological quality (HQ) in 13 sub-basins consisting of 49 sites with different degrees of conservation sampled between 2012 and 2016, *ii*) calculation of indicator values for 20 taxa to generate the proposed algorithm, and *iii*) validation of the algorithm in 26 new sites and calculation of its correlation with water quality and HQ. The ecological quality gradient to which the macroalgae responded was related to oxygen saturation, soluble reactive phosphorus, dissolved inorganic nitrogen, and hydromorphological quality. By assigning an indicator value to each group of species, we calculated the MEQI using the formula adapted from Pantle and Buck, indicating 5 different ecological quality classes. The MEQI index allowed us to establish references for good quality sites and altered sites with regards to chemical composition and structure. Therefore, the MEQI is recommended for biomonitoring in the Basin of Mexico and registering multiple environmental stressors.

Keywords: Aquatic biomonitoring; Ecological quality; Macroalgae; Montane rivers; Mexican Biotic Index

Resumen

Se evaluó la respuesta de las macroalgas a las diversas presiones antropogénicas en la cuenca de México para verificar si exhibían una correlación discriminatoria y explicativa para la evaluación ecológica y la propuesta del “Índice de Calidad Ecológica de las Macroalgas” (MEQI). La propuesta incluyó 3 pasos: *i*) caracterización del nivel

trófico y calidad hidromorfología (HQ) en 13 sub-cuencas con 49 sitios colectados entre 2012 y 2016 y que representan distintos estado de conservación, *ii*) cálculo del valor indicador de 20 taxones y la propuesta del algoritmo, y *iii*) validación del algoritmo en 26 nuevos sitios y su correlación con la calidad del agua y HQ. El gradiente de calidad ecológica al que respondieron las macroalgas estuvo relacionado con la saturación de oxígeno, el fósforo reactivo soluble, el nitrógeno inorgánico disuelto y la calidad hidromorfología. Al asignar un valor indicador a cada grupo de especies, calculamos el MEQI utilizando la fórmula adaptada de Pantle y Buck, indicando 5 clases de calidad ecológica diferentes. El MEQI ha establecido sitios de referencia de buena calidad y aquellos alterados en la composición química y estructura. Por tanto, se recomiendan para el biomonitoreo en la cuenca de México y el registro de sus múltiples estresores.

Palabras clave: Biomonitoreo acuático; Calidad ecológica; Macroalgas; Ríos de montaña; Índice biótico mexicano

Introduction

The Basin of Mexico has the potential to supply water to its nearly 22 million inhabitants through the many rivers that form in the mountainous systems surrounding it (Legorreta, 2009). However, continuous land use change and an increasing population have caused freshwater bodies to deteriorate in quantity and quality (Carmona-Jiménez & Caro-Borrero, 2017). This situation is partly due to a lack of regional environmental public policies that would improve the quality of the ecosystems through conservation practices involving local communities and other citizens under a governance approach. Despite their daily use of water, the inhabitants show little interest in the hydrological processes that relate to its supply. Unregulated practices of land use change, such as agriculture, tourism, urban pollution sources, and some anthropogenic activities related to structural modifications of the channels (dams, deviations, extractions, and confinements), promote the invasion and degradation of Protected Areas within Mexico City (Carmona-Jiménez & Caro-Borrero, 2017; Santos-Cerquera & Aguilar, 2016). To identify how these changes (natural and/or induced) derived from anthropogenic activity and affect biological communities, as well as how these communities respond to such changes, it is necessary to establish reference sites. According to Wallin et al. (2003), reference sites should have a set of conditions that most closely resemble the natural state, represented by a minimum extent of human intervention or no human intervention. In terms of biological quality, these conditions refer to the naturalness of the biotic structures and functions in the absence of significant human disturbances (Stoddard et al., 2006). In the montane rivers of the Basin of Mexico, these conditions are usually related to headwaters originating inside Protected Areas with the following characteristics: permanent water run-off with well oxygenated waters, low ionic content, and oligotrophic conditions (Carmona-Jiménez & Caro-Borrero, 2017). Moreover, eutrophication

and a decrease in biological diversity and ecological quality are some of the main effects of pollution. Studies of water quality in rivers are increasingly necessary. For example, Directive 2000/60 EC (European Commission, 2000) emphasized the need for water quality studies focused not only on physicochemical quality, but also the inclusion and monitoring of bioindicators as proxies for ecosystem quality (Poikane et al., 2016). In particular, conspicuous macroscopic algae growth (macroalgae, *sensu* Sheath & Cole, 1992) plays a preponderant ecological role in montane rivers, specifically in the cycling of nutrients and provision of food and refuge for freshwater fauna. Macroalgal communities have developed resilience strategies such as ecophysiological responses to environmental change, diverse asexual reproduction options, efficient dispersal and colonization mechanisms, and a wide variety of growth forms in fast current velocities (Bojorge-García et al., 2010; Borges & Necchi, 2006; Flor-Arnau et al., 2015). These communities have been proposed for quality assessments of aquatic ecosystems given that they respond morphologically and physiologically to physical environmental stressors (current velocity, changes in the flow regime, substratum diversity, water turbidity), and chemical changes in water (hardness, ionic content, pH, nutrient concentrations) (Poikane et al., 2016; Stancheva & Sheath, 2016). Likewise, soft-bodied algae are indicators of structural components associated with the ecological and functional integrity of ecosystems, among other things, due to their position at the base of the trophic chain and their physiological functions that contribute to the stability of the ecosystem—nitrogen-fixing heterocystous cyanobacteria (Rocha et al., 2017; Stancheva & Sheath, 2016).

The Basin of Mexico is home to a strongly adaptive community of macroalgae, most of which have temperate or boreal affinities (Rodríguez-Flores & Carmona-Jiménez, 2018). However, this resilient macroalgal community is threatened by a lack of efficient environmental policies and tools needed to evaluate and conserve the ecological quality

of Mexican aquatic systems, and their validation through bioindicator systems. Ecological quality is measured by the sum of its elements, between biological diversity and the ecological processes through which the ecosystem passes to reach its current state (Kaar, 1993). The evaluation of ecological quality and its relationship with conservation status and/or ecological deterioration requires the correct identification of possible biological indicators and their subsequent inclusion in indices that are easily applied and interpreted. To reduce the gap between academia and environmental policy, this study aims to analyze the macroalgal community within the Basin of Mexico and its relationship to water and hydromorphological quality

in order to develop the Macroalgae Ecological Quality Index (MEQI) as a potential management tool for water and ecological quality monitoring.

Materials and methods

The Basin of Mexico (Fig. 1) lies in the morphotectonic region of the Trans-Mexican Volcanic Belt (19°00'–19°40' N, 98°30'–99°30' W) and expands across 9,600 km², of which 5,518 km² are montane ranges that rise above 2,400 m asl. It was originally an endorheic basin, but was artificially opened to drain out excess water from storm peaks that caused flooding in Mexico City during the rainy

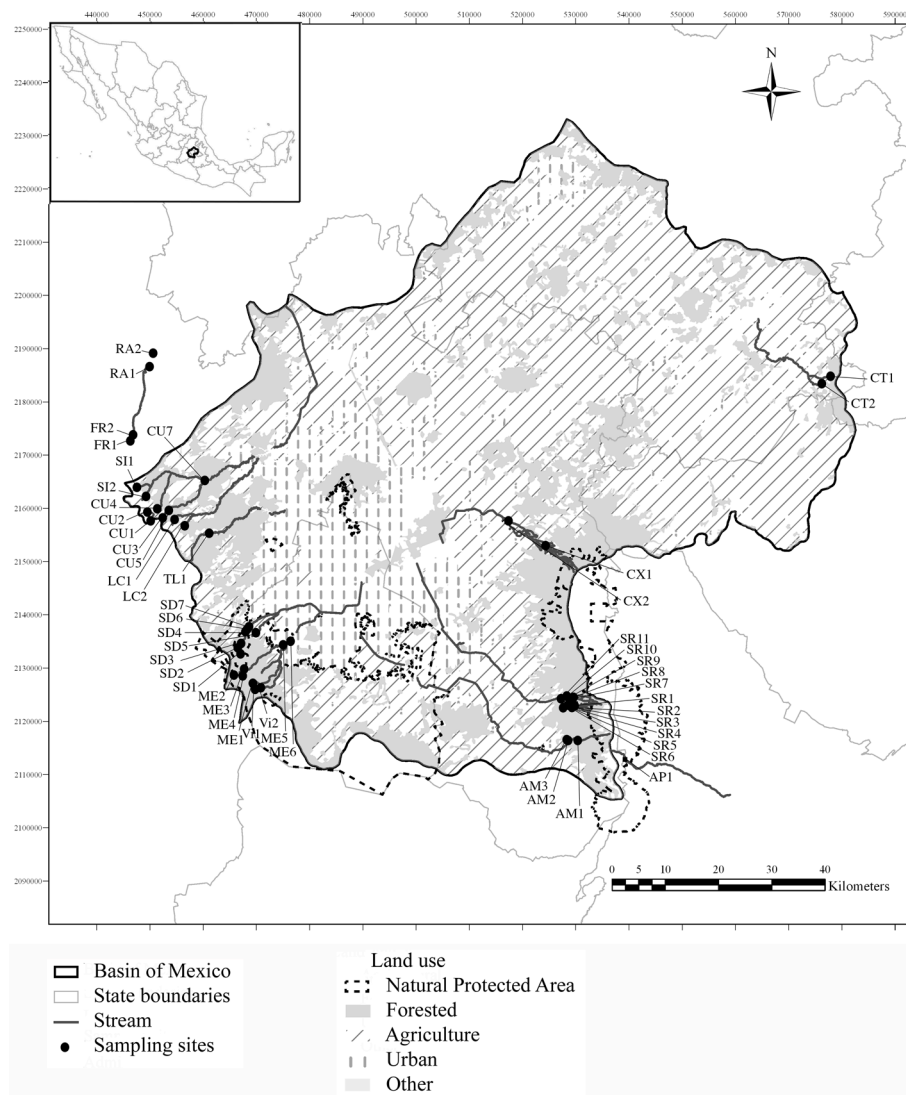


Figure 1. Map of the Basin of Mexico showing the 49 sampled points belonging to 13 rivers (sub-basins) and their land use. For sub-basin and site codes, see Table 1.

season (Legorreta, 2009). The region's climate is sub-humid and temperate (annual median temperature 13.4 °C, annual median precipitation between 1,200-1,500 mm), with abundant rain from June to October and a dry season from November to May (García, 2004). Its geological traits consist of andesitic to basaltic lavas (Ferrusquia-Villafranca, 1998). Forests of *Abies religiosa* (Kunth) Schltdl. et Cham., *Pinus hartwegii* Lindl., and *Quercus* spp. grow in the upper areas of the watershed, with mixed forests in the middle and lower areas (Ávila-Akerberg, 2010).

We selected 13 sub-basins each with a corresponding perennial river into the Basin of Mexico. Forty-nine locations were sampled within these rivers to represent the different conditions of conservation within forested areas. The criteria for site selection included the following: 1) streams that were reported as perennial in official documentation (cartographic information with hydrological network, GDF, 2012; INEGI, 2013); 2) location of headwaters within natural areas with minimal or no presence of urbanization and anthropogenic activities, with the objective of finding reference sites and conditions, and 3) the feasibility of collecting in middle and downstream sections to determine a gradient of change depending on level of anthropogenic intervention. In many cases, the river mouths were not sampled because they are piped and are part of the city's drainage system. Sampling was carried out between March 2012 and April 2016, with sites being visited several times (at least 2 and 3 times; Table 1) in contrasting seasons: rainy season (June to November) and dry season (December to May) (García, 2004). Water temperature, specific conductivity, and pH (Hanna Multiparameter probe 991300, Dallas, USA), oxygen saturation (YSI-85 meter, YSI, Ohio, USA), and current velocity (Global

Water FP111, Texas, USA) were recorded in situ. For each site, samples of 500 mL of water were filtered in situ and analyzed in the laboratory for selected nutrients according to official Mexican guidelines and international standards (DOF, 2003; Rice et al., 2017). Nitrite nitrogen ($\text{NO}_2\text{-N}$), nitrate nitrogen ($\text{NO}_3\text{-N}$), ammonium nitrogen ($\text{NH}_4\text{-N}$), dissolved inorganic nitrogen (DIN), and soluble reactive phosphorus (SRP, in theory mostly in the form of orthophosphate, $\text{PO}_4\text{-P}$) were analyzed with a DR 3900 laboratory spectrophotometer (Hach, Loveland, CO, USA; Hach, 2003). Trophic status was recognized as either oligo-mesotrophic (DIN, 0.7mg L^{-1} ; SRP, 0.25mg L^{-1}) or mesotrophic-eutrophic (DIN, 1.5mg L^{-1} ; SRP, 0.75mg L^{-1}) according to Dodds et al. (1998) and Dodds (2003).

Hydromorphological quality (HQ), referring to any modification of the flow regime, sediment transport, river morphology, or lateral channel mobility (Nardini et al., 2008), and anthropogenic effects were evaluated with reference to the Ecological Quality of Andean Rivers Index (Acosta et al., 2009). The HQ scale to classify the naturalness of the system —i.e., the conditions that present minimal alterations in riparian ecosystems of high montane streams— ranged from 24-120 points (> 96 high, 76-95 good, 51-75 moderate, 26-50 poor, < 25 bad) and considered the following aspects: 1) basin, including vegetation cover, bank stability, substrate diversity, and development of human activities surrounding the stream; 2) hydrology, evaluating the presence of dams, regimes of channel speed and depth, and alterations and state of the channel; and 3) anthropogenic disturbances, assessing the presence and impacts of human activity with direct effluents to the stream, urban development, human development, and presence of contamination.

Table 1

Physicochemical, hydromorphological characteristics, and macroalgal diversity of sampling sites in the Basin of Mexico (OS, oxygen saturation; T, water temperature; K_{25} , specific conductivity; CV, current velocity; SRP, soluble reactive phosphorus; DIN, dissolved inorganic nitrogen; HQ, hydromorphological quality).

River	Site and code <i>n</i> = number of collections	OS (%)	T (°C)	pH	K_{25} ($\mu\text{S cm}^{-1}$)	CV (m s^{-1})	SRP (mg L^{-1})	DIN (mg L^{-1})	HQ (pts.)	Taxa ^a
Coatlaco	Rancho nuevo I CT1 <i>n</i> = 1	86	11.2	7.4	61	0.23	0.38	0.16	100	3,9,11,12,15, 20
	Rancho nuevo II CT2 <i>n</i> = 1	53	13.5	6.8	68	0.26	1.20	0.10	82	10
Coaxacoaco	Santa Catarina CX1 <i>n</i> = 2	94	13.1	7.1	134	0.29	0.40	0.31	98	3,9,20
San Rafael	La Cabañita RA1 <i>n</i> = 2	100	13.9	6.8	80	0.52	0.21	0.57	114	10,20

Table 1. Continued

River	Site and code <i>n</i> = number of collections	OS (%)	T (°C)	pH	K ₂₅ (μS cm ⁻¹)	CV (m s ⁻¹)	SRP (mg L ⁻¹)	DIN (mg L ⁻¹)	HQ (pts.)	Taxa ^a
San Rafael- Tlalmanalco	La Planta RA2 <i>n</i> = 2	95	15.0	7.2	81	0.53	0.20	0.80	100	10
	San Rafael SR1 <i>n</i> = 2	93	12.0	5.5	72	0.51	0.61	0.13	85	5,11
	San Rafael canal 1 SR2 <i>n</i> = 2	100	9.9	7.5	138	0.51	0.48	0.14	118	3,5,9,10,11,13,20
	San Rafael canal 2 SR3 <i>n</i> = 2	89	11.3	6.8	136	0.90	0.51	0.96	94	3,5,9,10,11,13,20
	Agua dulce SR4 <i>n</i> = 2	90	9.0	6.6	138	0.57	0.80	0.48	100	3,20
	San Rafael canal 3 SR5 <i>n</i> = 2	97	11.6	7.0	137	2.05	0.33	0.86	96	5,11
	San Rafael bajo SR6 <i>n</i> = 2	97	11.6	7.0	137	0.46	0.33	0.86	96	3
	La Compañía SR7 <i>n</i> = 2	100	10.7	7.2	154	0.27	0.34	0.19	112	2,3,4,5,8,10,11
	Cosamala 1 SR8 <i>n</i> = 2	90	8.8	7.0	50	0.11	0.66	0.57	92	3,5,9,10
	La Compañía 2 SR9 <i>n</i> = 2	58	14.9	7.0	137	0.63	0.40	1.04	66	5,8,10
Ameca- Canal Nacional	Cosamala 3 SR10 <i>n</i> = 2	68	14.6	7.0	99	0.51	0.51	0.60	104	5,8,11
	Confluencia C-SR SR11 <i>n</i> = 2	58	14.9	7.0	137	0.63	0.40	1.04	66	10
	La Castañeda alto AM1 <i>n</i> = 2	90	8.3	7.3	255	0.54	0.27	0.76	120	5,10,14,15, 20
	La Castañeda AM2 <i>n</i> = 3	100	6.9	6.9	144	0.39	0.53	0.10	104	5,10,14,
	La Castañeda bajo AM3 <i>n</i> = 2	60	11.0	7.2	201	0.63	0.62	1.27	96	5,8,10,14,20
Frío	El Llano alto FR1 <i>n</i> = 1	100	15.6	7.2	35	0.85	0.48	0.49	102	4,10
	El Llano bajo FR2 <i>n</i> = 1	100	13.1	6.7	62	0.63	0.40	0.70	104	3,4,5,10
Viborillas	Monte Alegre I VI1 <i>n</i> = 2	129	11.6	7.4	50	0.16	0.21	0.47	82	3,9,11
	Monte Alegre II VI2 <i>n</i> = 2	89	17.7	7.2	63	0.04	0.23	1.23	77	11,16
Magdalena- Eslava	Manantial Eslava ME1 <i>n</i> = 2	43	8.8	5.7	112	0.20	0.19	0.54	110	3,13,15
	Chautitle alto ME2 <i>n</i> = 1	98	5.0	6.7	64	0.15	0.49	0.12	116	3,5,9,10,20
	Chautitle cañada ME3 <i>n</i> = 1	96	6.0	7.0	64	0.21	0.33	0.10	120	5,9,10,20

Table 1. Continued

River	Site and code <i>n</i> = number of collections	OS (%)	T (°C)	pH	K ₂₅ (μS cm ⁻¹)	CV (m s ⁻¹)	SRP (mg L ⁻¹)	DIN (mg L ⁻¹)	HQ (pts.)	Taxa ^a
Santo Desierto- Mixcoac	Truchero alto ME4 <i>n</i> = 2	82	7.6	7.2	63	0.20	0.28	0.24	88	5,10,11,20
	Confluencia E-M ME5 <i>n</i> = 2	50	12.0	7.1	86	0.98	0.39	2.78	54	4,9,10,19
	Santa Teresa ME6 <i>n</i> = 2	42	12.8	7.6	0	0.56	0.54	4.85	52	4
	Santa Rosa Alto SD1 <i>n</i> = 2	91	11.8	7.5	77	0.35	0.30	0.06	110	2,5,8,10,15
	Arroyo Desierto SD2 <i>n</i> = 2	90	9.6	7.0	89	0.17	0.50	0.14	120	11,20
	Santa Rosa Medio SD3 <i>n</i> = 1	80	10.0	6.5	79	0.86	0.50	0.10	114	4,9,10
	Convento SD4 <i>n</i> = 2	99	9.4	6.6	77	0.40	0.29	0.67	65	4,9,10
	La Capilla SD5 <i>n</i> = 1	80	13.0	7.3	150	0.60	0.26	0.11	104	4,14
	Escuela V-M SD6 <i>n</i> = 2	87	5.5	7.3	83	0.25	0.27	0.74	80	3,5,7,8,15, 20
Tlalnepantla	Truchero V-M SD7 <i>n</i> = 1	69	7.4	6.6	100	0.28	0.34	0.63	79	10
	San Luis Ayucan TL1 <i>n</i> = 2	55	10.8	6.6	60	0.90	0.97	1.71	68	10,18
La Colmena	La Caldera LC1 <i>n</i> = 1	100	8.1	6.7	35	0.05	0.53	0.04	116	10,13,15
	Xopachi LC2 <i>n</i> = 1	98	8.6	7.1	35	0.13	0.37	0.93	114	1,2,3,10,11,13
Cuautitlán	Los Organillos CU1 <i>n</i> = 3	88	10.6	6.5	101	0.20	0.50	0.68	104	3,11,13,16, 17,19
	Presa Iturbide CU2 <i>n</i> = 2	99	11.4	6.7	52	0.35	0.48	0.56	108	2,3,5,9,10, 11,14
	Manantial S-P CU3 <i>n</i> = 2	97	9.7	6.6	53	0.12	0.28	0.04	118	3,5,6
	Cortina P-I CU4 <i>n</i> = 1	80	13.2	6.5	53	0.64	0.96	1.55	88	11,13,15,18,19,20
	Capoxi CU5 <i>n</i> = 1	100	11.4	6.8	46	0.72	0.60	0.07	114	9,11,
	Manantial Capoxi CU6 <i>n</i> = 2	86	13.0	6.6	45	0.76	0.43	0.06	114	2,6,10
	Transfiguración CU7 <i>n</i> = 2	100	12.7	7.5	63	0.74	1.67	3.96	88	10,13,20
San Ildefonso	Las Palomas SI1 <i>n</i> = 2	100	11.8	7.0	49	0.16	0.23	0.43	120	5,11,14,19,20
	Truchero D.A. SI2 <i>n</i> = 2	100	10.3	7.2	54	0.67	0.37	0.26	107	5,10,13,15
Apatlaco	Apatlaco AP1 <i>n</i> = 1	94	10.7	6.9	58	0.98	0.25	0.16	111	9,20

^a: Number of taxa according to Table 3.

Each sampling site consisted of a randomly chosen 10 m long stream segment, which was divided into 5 equal quadrants where algae growths were collected and stream microhabitats were characterized (Necchi et al., 1995). We estimated a semi-quantitative abundance of macroscopic algae by measuring the percent coverage of each alga in each substrate through a visual estimation (Poikane et al., 2016; Rocha et al., 2017). The coverage was evaluated within a circular sampling unit 10 cm in radius (314 cm²), and each macroalgae growth was sampled in 5 units at different sampling points in the stream segment (Bojorge-García et al., 2010). The 5 samples of freshly collected material were reviewed under an optical microscope to identify characteristic reproductive or vegetative structures useful for species identification. Algae were identified to species level following taxonomic keys and specialized bibliographic resources (Anagnostidis & Komárek, 2005; Carmona-Jiménez & Necchi, 2002; Carmona-Jiménez & Vilaclara, 2007; Ettl & Gartner, 1988; Komárek, 2013; Rieth, 1980; Wher & Sheath, 2003; Zuccarello & Lokhorst, 2005). For taxonomic analyses, an Olympus BX51 microscope with a DP12 digital camera (Olympus, Tokyo, Japan) was used. Species with no reproductive structures or those that did not coincide with morphological characters described in the literature were not identified.

The MEQI was developed and validated in 3 steps. Step 1 “Establish the environmental differences between sampling sites”: first, we analyzed the environmental parameters (physicochemical variables and hydromorphological quality) separate from biological data as follows. Environmental factors were tested for collinearity to avoid correlation among variables. The mean values of variables selected for further analyses at each site were transformed through $\log(x+1)$. Principal Component Analyses (PCA) were performed to recognize the main environmental parameters that influence the ecological quality conditions of each site. Step 2 “The Macroalgae Ecological Quality Index (MEQI) proposal: calculate the indicator value”: Canonical Correspondence Analyses (CCA) followed by Montecarlo tests (999 permutation, $\alpha = 0.05$) were applied to determine the environmental preferences of macroalgal species by relating their distribution with physical, chemical, and hydromorphological parameters. The distribution of species scores along axis 1 and 2 of the CCA were used to separate species into 3 groups. The maximum and minimum values were used as a reference for tolerant and intolerant species, while intermediate values were used as a reference for detector species. Detector species are those that respond better to environmental changes and provide information for more than habitat (Tornés et al., 2007).

An indicator value was assigned to each group and used to calculate the MEQI using the formula adapted from Pantle and Buck (1955): $MEQI = \sum(v \cdot h) / \sum h$ where v is the species indicator value and h is the relative abundance (percent coverage) of the sample. All statistical analyses were performed in RStudio (R Studio Team, 2019) with the vegan (Oksanen et al., 2019) and ggplot2 (Wickham, 2016) packages. Step 3 “The MEQI validation”: we used a new test set to validate the MEQI proposed in the previous step. This test set consisted of 26 sites collected in the cool, dry season (February 13-21, 2019) spanning 3 of the largest sub-basins within the Basin of Mexico: Cuautitlán (8 sites), Santo Desierto (9 sites), and San Rafael-Tlalmanalco (9 sites). The HQ, physicochemical parameters, and macroscopic algae samples were obtained as previously described, and the MEQI was calculated for each site. To validate index performance in summarizing ecological quality, we calculated Pearson’s rank correlation coefficients of the MEQI with nutrient concentrations (DIN and SRP), oxygen saturation, and HQ. In this sense, an increase in MEQI scores reflected an increase in nutrient concentrations, organic pollution (OS as a proxy), and hydromorphological alterations.

Results

Streams within the Basin of Mexico showed relatively similar physicochemical characteristics (Table 1). The water temperature was temperate (5-17.7 °C), with variable oxygen saturation (42.5-100%), low to medium electrical conductivity (34.5-546 $\mu\text{S cm}^{-1}$), and slightly acidic to neutral pH (5.5-7.7). Current velocity ranged from slow to fast (0.05-2.05 m s⁻¹). Nutrient concentrations (SRP and DIN) were low in headwaters and increased towards mid and lower river reaches. The same pattern was observed in the hydromorphological quality assessment, where 29 sites were recognized as high quality, 14 sites as good quality, and 7 sites as moderate quality.

PCA results, considering nutrient concentrations (SRP and DIN), HQ, and physical parameters, showed that 61% of the variance in data among sampling sites was explained by the first principal component (Table 2). Groups I and II were both composed of sites with high or good HQ and OS, differing in nutrient loads as sites in group I had the lowest concentrations of SRP and sites in group II had the lowest concentrations of DIN. These 2 groups included headwaters (100%) and some sites from the streams’ middle courses (30%). Group III contained sites from middle (70%) and lower (100%) courses with the highest nutrient concentrations, lowest OS, and moderate HQ.

Table 2

PCA test considering inter-site variance in nutrient concentrations, hydromorphological quality and physical parameters in rivers of the Basin of Mexico (for abbreviation codes of the environmental variables and sites see Table 1).

	PC 1	PC 2	PC 3
Eigenvalue	2.5	1.3	1.2
% variance	30	15	15
Cumulative	30	46	61
Parameters	30	46	61
OS			0.49
K ₂₅		0.79	
SRP	0.30		
DIN	0.54		
HQ	-0.53		
CV		0.27	
Sites/group	III	II	I
CT1			0.64
CT2	1.67		
CX1		0.62	0.69
RA1			0.77
RA2			1.39
SR1			-2.52
SR2			1.38
SR3	0.68	0.96	
SR4		0.46	
SR5	1.21	1.99	1.64
SR6		0.76	
SR7		0.96	0.84
SR8	-0.32	-1.20	-0.58
SR9	2.27		
SR10	0.68	0.42	
SR11	2.27		
AM1		3.32	
AM2		0.66	
AM3		2.44	
FR1			2.20
FR2			0.20
VI1			1.76
VI2	1.09		1.00
ME1		1.37	
ME2	-2.11		

Table 2. Continued

	PC 1	PC 2	PC 3
ME3	-2.13		
ME4	-0.67	-0.12	-0.31
ME5	3.67		
ME6	5.12		
SD1			1.23
SD2	-1.37		
SD3		0.31	
SD4	0.15		
SD5		1.65	1.04
SD6		0.13	
SD7		0.62	
TL1	2.84		
LC1	-1.88		
LC2			0.13
CU1	-0.54	-0.10	-1.01
CU2	-0.74	-0.99	-0.18
CU3	-1.92		
CU4	1.71		
CU5			0.17
CU6	-0.50	-0.40	-0.20
CU7	3.60		
SI1			0.61
SI2			0.84
AP1			0.56

The MEQI proposal: the indicator value

Twenty macroalgal species were collected and identified (Table 3). Species with wide distribution across the basin (> 40% of the sites) and the greatest cover (> 23%) were *Nostoc* aff. *parmelioideus*, *Placoma regularis*, *Prasiola mexicana*, *Spirogyra* sp., and *Vaucheria bursata*. Other species high in percent coverage but restricted to certain sites included *Microcoleus* aff. *autumnale*, *Draparnaldia mutabilis*, *Tetraspora gelatinosa*, *Stigeoclonium tenue*, and *Batrachospermum gelatinosum*.

CCA revealed that most sampled streams had low to moderate nutrients, as eutrophic conditions were less frequent (Table 4, Fig. 2). The main ordination of the sampling sites related to environmental parameters was verified along axis one of the CCA, with 42.7% of the variance explained. The canonical coefficients revealed DIN ($r = 0.65$) to be the most important environmental

Table 3

Macroalgal diversity in rivers of the Basin of Mexico. * Species registered in the collection to validate the MEQI index.

Taxa	Frequency (number of sites collected)	Cover percentage (min-max%)
Cyanobacteria		
1. <i>Calothrix</i> sp.	1	1
2. <i>Coleodesmium wrangelii</i> Borzi ex Geitler	5	1-11
3. <i>Nostoc</i> aff. <i>parmelioides</i> Kützing ex Bornet & Flahault *	24	1-16
4. <i>Microcoleus</i> aff. <i>autumnale</i> (Gomont) Strunecky, Komárek & Johansen in Strunecky et al. *	9	1-26
5. <i>Placoma regulare</i> Brody & Ingerfeld *	27	1-26
Rhodophyta		
6. <i>Batrachospermum</i> <i>gelatinosum</i> (Linnaeus) De Candolle *	4	2-15
7. <i>Paralemanea mexicana</i> (Kützing) Vis & Sheath *	1	3
8. 'Chantransia' stage of undefined Rhodophyta *	6	1-5
Chlorophyta		
9. <i>Ulothrix</i> sp.	15	0.1-8
10. <i>Prasiola mexicana</i> Agardh *	41	0.5-76
11. <i>Spirogyra</i> sp. *	20	1-23
12. <i>Mougeotia</i> sp.	1	1
13. <i>Cladophora</i> sp. *	8	1-5
14. <i>Rhizoclonium</i> sp.	7	1-6
15. <i>Oedogonium</i> sp. *	10	1-45
16. <i>Tetraspora gelatinosa</i> (Vaucher) Desvaux	5	2-5
17. <i>Draparnaldia mutabilis</i> (Roth) Bory	1	2
18. <i>Stigeoclonium tenue</i> (Agardh) Kützing *	3	1-15
Xanthophyta		
19. <i>Tribonema vulgare</i> Pascher.	4	1-4
20. <i>Vaucheria bursata</i> (Müller) Agardh *	26	1-23

Table 4

The CCA analysis and inter-set coefficients of the canonical correlation of the environmental variables and the macroalgal indicator value (v) in rivers of the Basin of Mexico (for abbreviation codes of the environmental variables see Table 1).

	CCA1	CCA2	
Eigenvalue	0.61	0.35	
p	0.002	0.007	
% variance	42.7	24.9	
Cumulative	42.7	67.6	
Parameters			
OS	-0.37		
SRP		-0.20	
DIN	0.65		
HQ	-0.50		
CV		-0.39	
Species			v
<i>Coleodesmium wrangelii</i>	-1.11	-0.69	1
<i>Ulothrix</i> sp.	-0.40	-0.85	1
<i>Nostoc</i> aff. <i>parmelioides</i>	-0.34	0.50	1
<i>Vaucheria bursata</i>	-0.32	-0.25	1
<i>Prasiola mexicana</i>	-0.31	0.16	1
<i>Batrachospermum</i> <i>gelatinosum</i>	-0.31	0.76	1
<i>Placoma regulare</i>	-0.15	-1.10	3
<i>Cladophora</i> sp.	-0.13	-0.31	3
<i>Spirogyra</i> sp.	-0.12	1.26	3
<i>Rhizoclonium</i> sp.	-0.07	-0.23	3
<i>Oedogonium</i> sp.	0.35	-0.38	3
<i>Tetraspora</i> sp.	0.38	4.57	3
'Chantransia' stage	0.60	-0.95	5
<i>Stigeoclonium tenue</i>	0.95	-2.53	5
<i>Tribonema vulgare</i>	2.88	-0.91	5
<i>Microcoleus</i> aff. <i>autumnale</i>	4.27	-0.19	5

variable to the ordination correlated with the positive quadrant, while HQ (r = -0.50) and OS (r = -0.37) were the most important environmental variables to the ordination strongly correlated with the negative quadrant. Axis 2 (24.9% of explained variance) was negatively correlated with current velocity (r = -0.39) and SRP (r = -0.20). Indicator values were assigned following Signini (2002), who recommends normalizing the values with other

aquatic bioindicators. Species below the 25th percentile (≤ -0.31) were considered indicators of good ecological status ($v = 1$), while species above the 75th percentile (≥ 0.54) were considered indicators of poor ecological status ($v = 5$). Species with scores between the 25th and 75th percentile regions (> -0.31 and < 0.54), or in the transition zone, were considered tolerant species ($v = 3$).

Species' percent coverage and indicator values were used to calculate the MEQI applying the formula adapted from Pantle and Buck (1955). The index values ranged from 1 to 5, representing 5 different ecological qualities determined by macroalgal representatives (Table 5) and those established by the Water Framework Directive (2000) (European Commission, 2000): high (1.0-1.5), good (1.51-2.5), moderate (2.51-3.5), poor (3.51-4.5), or bad (4.51-5.0) ecological quality. Values lower than 1.5 can be considered as reference sites with high ecological quality, and values higher than 3.51 indicate disturbed sites. Intermediate classes indicate species that can tolerate high concentrations of DIN and/or SRP, low OS, or disturbances in the HQ. A recurrent association of *N. aff. parmelioideis*, *C. wrangelii*, *P. regulare*, and *V. bursata* was recognized in sites scored as having high and good ecological quality. Associations with *P. mexicana*, *Cladophora*, and *Spirogyra* were found in good and moderate scored sites. It should be noted that *Prasiola* was found in sites ranging from high to poor ecological quality, always in association with high water flow. *M. aff. autumnale* and *S. tenue* were found in association with sites of poor and bad ecological quality.

MEQI validation

A total of 12 macroalgae species were recorded in the validation collection (Table 3). In general, the MEQI was consistent with the value of hydromorphological quality and trophic status. Headwaters presented good and high scores, while an evident decrease in hydromorphological quality was related to poor and bad MEQI classes (Fig. 3, Table 6). We found a strong positive correlation between the MEQI scores and both DIN and SRP, and a moderately positive correlation with HQ ($p < 0.05$ for all correlations) (Table 7). Modifications in canal structure and the extraction and retention of water in the rivers confirmed the moderate HQ value in the middle and lower portions of the 3 sub-basins.

Discussion

The most important causes of environmental degradation in these montane rivers have been the alteration of the riverbed structure and the presence of hydraulic infrastructure such as dams, diversion channels, and pipelining of springs, leading to the loss of habitat heterogeneity and/or physical fragmentation. These alterations have been historically overlooked in the region, with intensive extraction of surface run-off and groundwater without prior ecological knowledge or any governmental regulation to maintain an ecological equilibrium (Caro-Borrero et al., 2017; Sánchez, 2007). In general, the highest nutrient concentrations (evaluated as

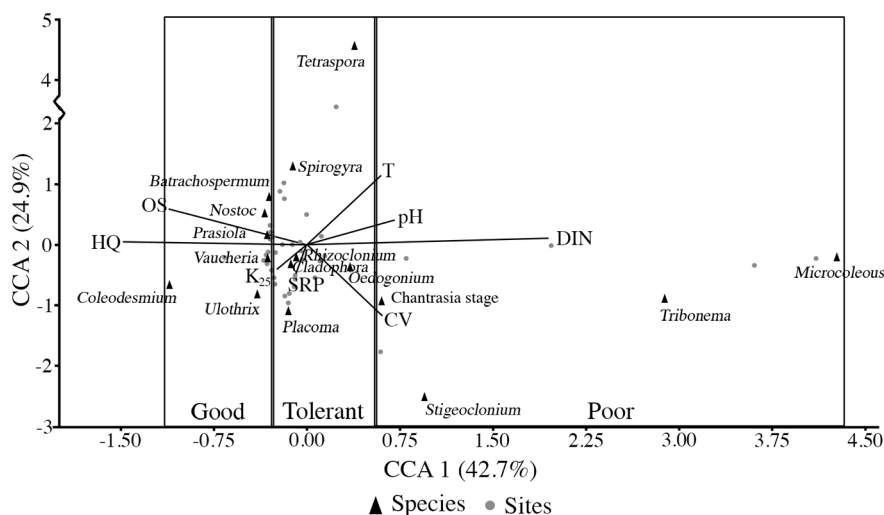


Figure 2. Scatterplots based on the CCA of the macroalgae samples (percentage cover) from 49 sites in the Basin of Mexico, correlated with the environmental variables to the ordination of axes 1 and 2. Ordination of the macroalgae taxa according to their ecological optima, recognizing indicators of good and poor ecological status and tolerant species (for codes of the environmental variables, see Table 1).

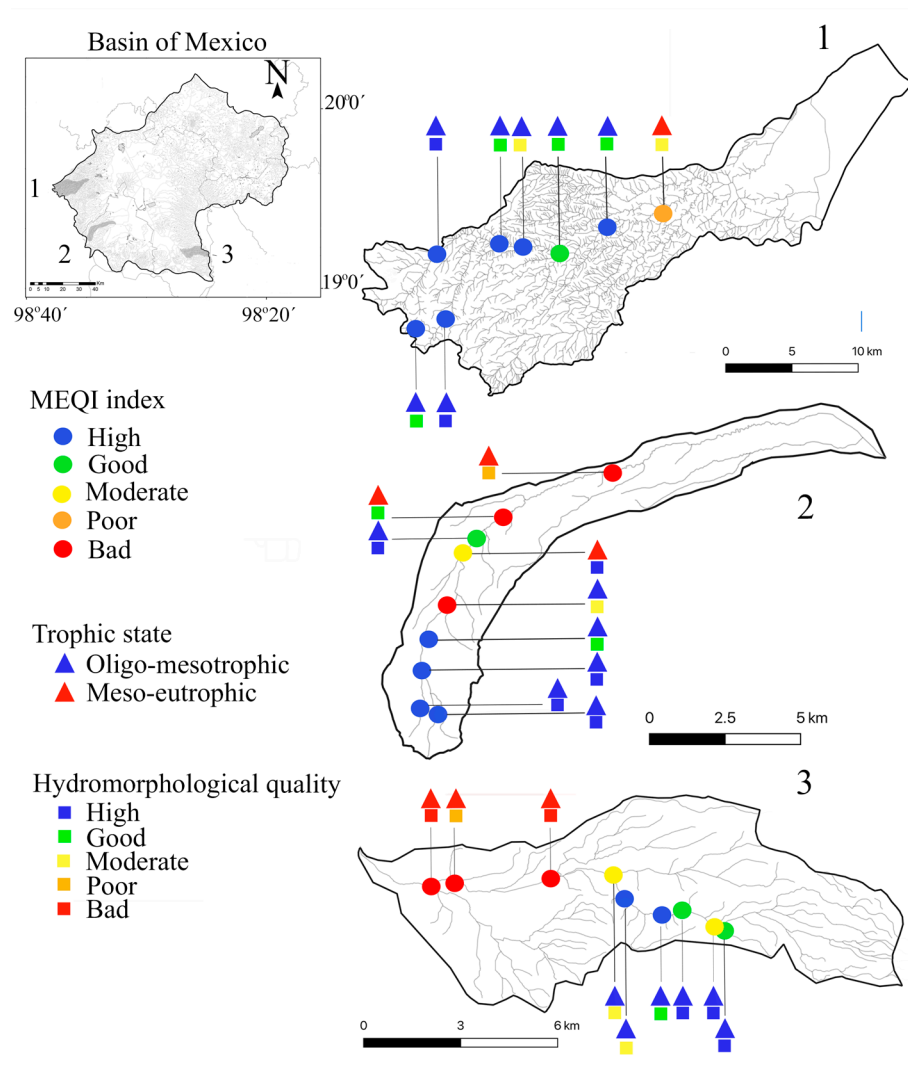


Figure 3. The MEQI score, ecological quality, and physicochemical quality in 1, Cuautitlán; 2, Santo Desierto-Mixcoac, and 3, San Rafael-Tlalmanalco sub-basins of the Basin of Mexico.

SRP and DIN) and the lowest OS were observed mainly in the lower reaches of the rivers. This condition highlights the increase in degradation of ecological quality, which leads to the loss of habitat and ecosystem services, and in some cases may even represent a risk to human health (Bonilla-Lemus et al., 2014; Mazari-Hiriart et al., 2014).

The greater macroalgal richness and diversity found in most headwaters of sampled rivers, including for example *C. wrangelii*, *N. aff. parmelioideis*, *P. regularis*, *B. gelatinosum*, *Paralemanea mexicana*, *V. bursata*, and *P. mexicana*, were related to potential reference conditions described in the region by other studies (Carmona-Jiménez & Caro-Borrero, 2017; Rodríguez-Flores & Carmona-Jiménez, 2018).

The macroalgal diversity was represented by a resilient community typically associated with montane rivers of central Mexico (above 2,300 m). Affinities of macroalgal communities between the Basin of Mexico and temperate eastern regions of North America indicate a possible biogeographic connection with the Nearctic region under similar environmental requirements (Sheath & Cole, 1992).

Our results suggest that the specimens studied tend to exhibit dispersal patterns similar to those described previously (Branco et al., 2014). Species of Cyanobacteria, Chlorophyta, and Xanthophyceae (Heterokontophyta) have a broad range of different ecological requirements that determine their dominance in many aquatic ecosystems. This ability results in a wider ecological and geographic

Table 5

The MEQI value and representative of the ecological quality of the assemblages (EQ) from 50 sites in the Basin of Mexico (for abbreviation codes of the sites see Table 1).

MEQI	Assemblages representative	Sites and MEQI
High (1.0-1.5). At least 75% of the assemblage are indicator species of high EQ	i) <i>Nostoc</i> aff. <i>parmelioides</i> - <i>Spirogyra</i> sp.- <i>Vaucheria bursata</i> - <i>Cladophora</i> sp. ii) <i>N. aff. parmelioides</i> - <i>Coleodesmium wrangelii</i> - <i>Calothrix</i> sp.- <i>Placoma regulare</i> ii) <i>N. aff. parmelioides</i> - <i>Draparnaldia mutabilis</i> - <i>Batrachospermum gelatinosum</i> iv) <i>Tetraspora vulgaris</i> - <i>Spirogyra</i> sp.	AP1 (1), CT2 (1), CU6 (1.0), CU8 (1.0), CX1 (1.0), RA1 (1.0), RA2 (1.0), SD7 (1.0), SR11 (1.0), SR4 (1.0), SR6 (1.0), CT1 (1.12), AM2 (1.13), ME3 (1.24), SR8 (1.27), CU1 (1.29), ME1 (1.32), SD2 (1.33), ME2 (1.35), FR1 (1.36), FR2 (1.39), LC1 (1.41), SD1 (1.42), SD3 (1.43), SD4 (1.42), SR7 (1.45), AM3 1.45, SR9 (1.45)
Good (1.51-2.5) At least 25% of the assemblage are indicator species of high EQ	i) <i>N. aff. parmelioides</i> - <i>Ulothrix</i> sp.- <i>Oedogonium</i> sp.- <i>V. bursata</i> ii) <i>P. regulare</i> - <i>Prasiola mexicana</i> - 'Chantransia' stage	SD6 (1.54), SR2 (1.6), CU5 (1.66), CU7 (1.66), ME4 (1.83), VI1 (1.84), SR3 (1.9), CU4 (1.98), CU2 (1.99), AM1 (2.45)
Moderate (2.51-3.5) At least 50% of the assemblage are tolerant species	i) <i>P. mexicana</i> - <i>Paralemanea mexicana</i> - <i>P. regulare</i> ii) <i>P. mexicana</i> - <i>Microcoleus</i> aff. <i>autumnale</i> iii) <i>P. mexicana</i> - 'Chantransia' stage- <i>Oedogonium</i> sp. iv) <i>Rhizoclonium</i> sp.- <i>M. aff. autumnale</i> - <i>Cladophora</i> sp. v) <i>P. mexicana</i> - <i>Spirogyra</i> sp.	SI2 (2.58), LC2 (2.73), SI1 (2.91), SR1 (3.0), SR5 (3.0), VI2 (3.0), SD5 (3.4), ME5 (3.0)
Poor (3.51-4.5) At least 25% of the assemblage are indicator species of poor EQ	i) <i>P. mexicana</i> - <i>Stigeoclonium tenue</i> - <i>M. aff. autumnale</i>	TL1 (3.67), ME6 (4.44)
Bad (4.51-5.0) At least 75% of the assemblage are indicator species of bad EQ	i) <i>M. aff. autumnale</i> - <i>S. tenue</i>	ME5 (4.64)

distribution (Bojorge-García et al., 2010; Sheath & Cole, 1992). Thus, we observed that Cyanobacteria and Chlorophyta species with higher dispersal abilities —i.e., spores and resistant vegetative fragments, which facilitate dispersion and desiccation tolerance— are affected by local environmental variables —e.g., canopy cover—, whereas Rhodophyta (*B. gelatinosum* and *Paralemanea mexicana*), which have low dispersal abilities and long life cycles, are apparently influenced by local variables —e.g., temperature and current velocity— and reflect the conditions of long-term water chemistry (Stacheva & Sheath, 2016).

Although some macroalgae species such as *P. regulare* and *P. mexicana* were observed in the middle and lower river reaches, the deleterious effects of anthropogenic pressures were apparent in terms of percent coverage, richness, and species diversity. The HQ was of special relevance since it was found to be significantly correlated with nutrient

enrichment and the changes associated with the riparian zone, substratum diversity, channel fragmentation, and water retention, all properties related to processes that favor nutrient enrichment and accumulation. Additionally, morphological and functional features of algal populations and their relationship to the aquatic environment were recognized.

Cyanobacteria species such as *N. aff. parmelioides* and *C. wrangelii* can fix atmospheric nitrogen under limiting conditions and form important percent coverage in long segments of Central Mexican rivers (Bojorge-García et al., 2010; Rodríguez-Flores & Carmona-Jiménez, 2018). Stacheva and Sheath (2016) point out that the abundance of heterocystous cyanobacteria such as *Nostoc* and *Coleodesmium* are a promising metric for bioevaluations, since they are consistently reported as being negatively correlated with the amount of dissolved nitrogen in the water. Some species previously reported as highly

Table 6

Physicochemical, hydromorphological characteristics and MEQI value of validation sampling sites in the Basin of Mexico (for abbreviation codes of the environmental variables see Table 1).

Sub-basin, sites and altitude (m asl)	K ₂₅ (μS cm ⁻¹)	OS (%)	SRP (mg L ⁻¹)	DIN (mg L ⁻¹)	HQ (pts)	Macroalgae (Cover percentage)	MEQI
San Rafael-Tlalmanalco							
Waterfall Diamantes 3266	174	8.1	0.45	0.05	120	<i>Vaucheria bursata</i> (20) <i>Cladophora</i> sp. (15)	1.8
Waterfall Negro 3094	161	8.4	0.31	0.04	120	<i>Placoma regulare</i> (10)	3
Vereda 2956	151	8.1	0.31	0.04	102	<i>P. regulare</i> (10) <i>P. mexicana</i> (30) <i>V. bursata</i> (30) <i>Cladophora</i> sp. (20)	1.6
Canal 2915	142	7.9	0.38	0.04	95	<i>Nostoc</i> aff. <i>parmelioides</i> (30) <i>P. regulare</i> (5) <i>P. mexicana</i> (10) <i>V. bursata</i> (15)	1.16
Agua dulce 2788	151	8.2	0.33	0.09	63	<i>P. regulare</i> (1) <i>P. mexicana</i> (15)	1.12
Dos aguas 2676	123	7.7	0.34	0.12	55	<i>P. mexicana</i> (5)	1
San Rafael 2525	261	5.8	2.58	21.52	18	<i>Stigeoclonium tenue</i> (20)	5
UAM 2429	415	3.0	3.24	15.4	37	<i>S. tenue</i> (80)	5
Tlalmanalco 2400	677	2.7	2.52	52.06	24	<i>S. tenue</i> (15)	5
Cuautitlán							
Spring 3379	39	7.6	0.15	0.13	111	<i>N. aff. parmelioides</i> (5)	1
Spring Iturbide 3317	60	7.7	0.14	0.23	72	<i>N. aff. parmelioides</i> (27) <i>P. regulare</i> (10) <i>P. mexicana</i> (20) <i>V. bursata</i> (15)	1.27
Iturbide 3205	59	8.4	0.13	0.16	104	<i>N. aff. parmelioides</i> (10)	1
El paraíso 2703	51	8.0	0.36	0.21	120	<i>P. regulare</i> (25) <i>P. mexicana</i> (100) <i>V. bursata</i> (30)	1.34
El polvorín 2612	70	7.2	0.21	0.21	76	<i>P. mexicana</i> (90) <i>V. bursata</i> (50)	1
Cantera 2578	72	7.5	0.15	0.38	65	<i>P. regulare</i> (55) <i>P. mexicana</i> (20) <i>V. bursata</i> (27.5)	2.07
Transfiguración 2515	70	7.6	0.29	0.23	92	<i>P. regulare</i> (15) <i>P. mexicana</i> (100)	1.26
Yondese 2396	113	7.6	0.42	0.32	85	<i>P. regulare</i> (10) <i>V. bursata</i> (30) <i>Cladophora</i> sp. (25)	2.07

Table 6. Continued

Sub-basin, sites and altitude (m asl)	K ₂₅ (μS cm ⁻¹)	OS (%)	SRP (mg L ⁻¹)	DIN (mg L ⁻¹)	HQ (pts)	Macroalgae (Cover percentage)	MEQI
Nicolás Romero 2333	175	6.8	1.75	1.94	53	<i>S. tenue</i> (40)	5
Santo desierto							
San Miguel 3310	57	7.7	0.26	0.06	117	<i>N. aff. parmelioide</i> s (20) <i>P. regulare</i> (1)	1.09
Conejos 3314	72	7.4	0.13	0.11	111	<i>N. aff. parmelioide</i> s (20) <i>P. regulare</i> (1)	1
Confluencia alta 3123	75	7.5	0.17	0.15	108	<i>V. bursata</i> (10)	1
Camino 2840	82	8.2	0.14	0.42	93	<i>V. bursata</i> (7.5)	1
Convento 2907	115	5.7	0.36	0.29	66	‘Chantransia’ stage (20)	3
Escuela 2702	110	7.79	0.14	0.24	102	<i>N. aff. parmelioide</i> s (30) <i>P. regulare</i> (20) <i>P. mexicana</i> (15)	1.61
Truchero 2734	130	7.8	0.08	0.24	91	<i>P. mexicana</i> (15)	1
Cancha 2719	369	5.7	2.48	2.12	46	without algae	5
Guadalupe Victoria 2491	347	1.6	2.49	30.04	37	<i>S. tenue</i> 10	5

tolerant to ecological degradation (Loza et al., 2013; Pawlik-Skowrońska, 2003), such as *M. aff. autumnalis* and *Stigeoclonium tenue*, were associated with low HQ, decreased water flow, and increased nutrients. Algae possess high nutrient-binding capacities due to the presence of polysaccharides, proteins, or lipids on the surface of their cell walls, which contain functional groups like aminos, hydroxyls, carboxyls and sulfates, and can act as binding

sites for nutrients and even metals (Mehta & Gaur, 2005). Representative species of the good HQ species group, as well as an increase in nutrients, are generally observed in protected natural areas with considerable recreational activity. In particular, *P. regulare*, *V. bursata*, *Ulothrix* sp., *Spirogyra* sp., *Cladophora* sp., and *P. mexicana* are common inhabitants of oligotrophic to mesotrophic peri-urban rivers in the Trans-Mexican Volcanic Belt (Bojorge-García et al., 2010).

Macroalgae are seldom used on their own in stream bioassessments, but are grouped with phytobenthos or macrophyte systems (Stancheva & Sheath, 2016), with only a few exceptions (Schneider & Lindström, 2009, 2011). In the montane rivers of the Basin of Mexico, benthic macroalgae far outnumber aquatic macrophytes, and ecomorphological responses to spatial and temporal changes are short term. Since biological indices aim to assess specific situations based on local conditions and local biodiversity, we propose the use of the MEQI to assess the ecological quality in this particular region on the basis of its own macroalgal flora. Numerous species of algae were present in all surveyed rivers, and many of them may be

Table 7

Spearman’s rank correlations between the MEQI and physicochemical and hydromorphological characteristics of validation sampling sites in the Basin of Mexico (for abbreviation codes of the environmental variables see Table 1).

Coefficient	K ₂₅	OS	HQ	DIN	SRP
<i>r</i> Pearson	0.79	-0.80	-0.77	0.75	0.94
R ²	0.62	0.64	0.59	0.56	0.87
<i>p</i> -value	1.37 E ⁻⁵	8.15 E ⁻⁶	3.12 E ⁻⁵	6.56E -5	1.84 E ⁻¹⁰

regarded as reference species for the ecological quality of high montane rivers in the Trans-Mexican Volcanic Belt. The set of species presented here should be considered in future biomonitoring protocols, with the aim of reaching regional agreements and public policy implementations that avoid further degradation of the peri-urban rivers in the Basin of Mexico. The great morphological variation within *Microcoleus*, *Nostoc*, and *Spirogyra* populations suggests the existence of undescribed species whose taxonomic identity may be confirmed with molecular tools and the assessment of habitat preferences. Ecological quality is a holistic concept, not confined to a single taxonomic group, but rather requiring the understanding of different biological groups and their interactions within a particular catchment area (Kelly, 2013); different groups respond to different pressures, allowing us to evaluate ecological quality in an integrative way. These bioassessment methods are intended to detect impairment rather than to diagnose cause, since many pressures act in a combined, not isolated way. Oftentimes it is not easy to establish a correlation between a particular pressure and its causal relationship (Poikane et al., 2016), and future work to establish potential mechanisms that more precisely link river conditions will facilitate appropriate restorative actions.

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