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Ecology

Functional reef fish groups of the Mexican Caribbean: implications of habitat complexity

Grupos funcionales de peces de arrecife del Caribe mexicano: implicaciones de la complejidad del hábitat

Rosalía Aguilar-Medrano a, *, Jesús Ernesto Arias-González b

*Corresponding author: liabiol@gmail.com (R. Aguilar-Medrano)

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Abstract

The Caribbean Sea is considered a unique biogeographic region and is among the top 5 world hotspots of marine biodiversity. Yet, over the last years it is among the regions where the most serious declines in terms of reef area, health, and productivity is occurring. In this study, the coral reef fish community of the Mexican Caribbean was analyzed and classified into functional groups. The aims of this study were: 1) determine the functional relationships of the fish community based on diet and morphology, and 2) determine the extent of the influence of habitat complexity and the protection of the Sian-Ka'an reserve on the fish richness and functional diversity of the ecosystem. Eleven reef localities of the Yucatán Peninsula and 160 species were analyzed. Twenty-eight functional groups were described for both ecological and morphological variables. Our results indicate that variation in habitat complexity was a significant factor related to fish richness and functional diversity; however, the protection of the reserve does not show a relationship with any of the response variables. On the other hand, the slope is the reef zone with the highest fish richness in all localities, creating a belt of richness and connectivity in the whole area.

Keywords: Reef fish; Functional groups; Yucatán Peninsula; Habitat complexity; Marine reserve

Resumen

El mar Caribe es considerado una región biogeográfica única y entre los 5 principales hotspots de biodiversidad marina. Sin embargo, en los últimos años es de las regiones con las más graves disminuciones en términos de área arrecifal, salud y productividad. En este estudio, la comunidad íctica de los arrecifes del Caribe mexicano fue analizada y clasificada en grupos funcionales. Los objetivos de este estudio fueron: 1) determinar las relaciones funcionales

^a Laboratorio de Taxonomía y Ecología de Peces, Centro de Investigación y de Estudios Avanzados, IPN, Unidad Mérida, Antigua carretera a Progreso Km 6, 97310 Mérida, Yucatán, Mexico

^b Laboratorio de Ecología de Ecosistemas de Arrecifes Coralinos, Departamento de Recursos del Mar, Centro de Investigación y de Estudios Avanzados del Instituto Politécnico Nacional, Unidad Mérida, Antigua carretera a Progreso Km 6, Apartado Postal 73 CORDEMEX, 97310 Mérida, Yucatán, Mexico

de la comunidad de peces con base en la dieta y la morfología, y 2) determinar la influencia de la complejidad del hábitat y la protección de la reserva Sian-Ka'an en la riqueza de peces y la diversidad funcional del ecosistema. Se analizaron 11 arrecifes de la península de Yucatán y 160 especies. Mediante variables ecológicas y morfológicas, se describieron 28 grupos funcionales. Nuestros resultados indican que la complejidad del hábitat es un factor relacionado con la riqueza de peces y la diversidad funcional. La protección de la reserva no muestra relación con ninguna de las variables de respuesta. Por otro lado, la pendiente es la zona de arrecife con mayor riqueza de peces en todas las localidades, creando un cinturón de riqueza y conectividad en toda la zona.

Palabras clave: Peces de arrecife; Grupos funcionales; Península de Yucatán; Complejidad del hábitat; Reserva marina

Introduction

The diversity of coral reef communities has always been considered to be higher than other marine habitats (Emery, 1978). Although coral reef ecosystems cover less than 0.07% of the ocean surface, they are the home to 25% of all known marine species on the planet (Spalding & Brown, 2015; Spalding & Grenfell, 1997; Spalding et al., 2001). Calculations of the fish diversity in these ecosystems propose approximately 4,000 to 8,000 species, indicating that coral reef ecosystems are the host of one of the most diverse fish assemblages to be found anywhere on earth (Lieske & Myers, 2001; Spalding et al., 2001).

The Caribbean Sea is considered a unique biogeographic region and is among the top 5 world hotspots for marine biodiversity (Rivera-Monroy et al., 2004). It constitutes 12% of the total reefs of the world, shelters approximately 500 species of fishes, and presents the second largest reef barrier in the world, the Mesoamerican Barrier Reef System (MBRS) (Cortés, 2003; Hughes, 1994). Yet, over the last years, the Caribbean Sea is among the regions where the most serious declines in terms of reef area, health, and fisheries productivity is occurring (Gardner et al., 2003; Miloslavich & Klein, 2005).

Coral reefs around the world are under stress by the combined effect of negative factors. As a result of global climate change, we are facing issues such as ocean warming (Frieler et al., 2012; Roxy et al., 2016; Spalding & Brown, 2015) and acidification (Wisshak et al., 2012). Invasive species such as the lionfish (*Pterois volitans*) and seagrass species (*Halophile stupulacea*) are dramatically changing the trophic webs (Arias-González et al., 2011; Williams, 2007), and anthropogenic effects such as overfishing and pollution have severely degraded these ecosystems (Sandin et al., 2008).

Although coral reefs are in crisis worldwide (Bellwood et al., 2004; De'ath et al., 2012; Huang, 2012; Nyström et al., 2000), the response to stressors will vary depending on the resilience of these ecosystems (Roff & Mumby, 2012). In ecology, resilience is the capacity of an ecosystem to respond to a perturbation or disturbance by resisting

damage and recovering (Folke et al., 2004). The analyses of functional groups, born as a way to measure the trophic web diversity of ecosystems (Petchey & Gaston, 2006), has shown that the level of redundancy within these functional groups can be a measure of reef assemblage stability, and ultimately a measure of coral reef resilience to perturbations (Clarke & Warwick, 1998; Nyström, 2006).

Functional groups are collections of species that perform a similar ecological function in the ecosystem, irrespective of their taxonomic affinities (Bellwood et al., 2004; Steneck, 2001; Steneck & Dethier, 1994). The ability of species to use specific resources in specific areas of the ecosystem relies on its morphological design (Fulton, 2007; Wainwright & Bellwood, 2002). Through the study of these strong relationships between habitat use, morphology, and diet on coral reef fishes, it is possible to recognize groups with highly specific niche and consequently, highly specific functions in the ecosystem (Aguilar-Medrano & Calderón-Aguilera, 2015; Aguilar-Medrano et al., 2011, 2012; Fulton, 2007; Wainwright & Bellwood, 2002).

Morphology has been used to aid in determining functional groups (Córdova-Tapia & Zambrano, 2016; Villéger et al., 2010), conditioning morphology to a number of measurements that we cannot readily visualize as descriptors of shape differences. Aguilar-Medrano and Calderón-Aguilera (2015) proposed a classification of functional groups based on diet and morphology, using geometric morphometrics to define shape variations. Geometric morphometric analyses not only solves many of the problems confronting traditional methods of measurement by offering precise and accurate descriptions, but also it serves the equally important purposes of visualization, interpretation, and communication of results (Zelditch et al., 2004). Thus, in order to study the relationship between fish functional groups and habitat complexity along the coast of the Yucatán Peninsula we have used the classification of Aguilar-Medrano and Calderón-Aguilera (2015).

For a long time researchers have investigated the relationship between fish species richness-diversity and

habitat complexity variables such as topography, depth, substrate, slope, currents, and coral cover, among others (Aguilar-Medrano & Calderón-Aguilera, 2015; Arias-González et al., 2011; Carpenter et al., 1981; Dominici-Arosemena et al., 2005; Gosline, 1965; Harmelin-Vivien, 1977; Levin, 1993; Luckhurst & Luckhurst, 1978; McGehee, 1994; Roberts & Ormond, 1987; Williams, 1982). All variables have shown differences according to the region; however, habitat complexity and anthropogenic pressure appear to be the factors with the highest correlation to fish abundance and diversity (Arias-González et al., 2008, 2011; Bell & Galzin, 1984; De'ath et al., 2012; Huang, 2012; McCormick, 1994; Nemeth & Appeldoorn, 2009).

In the present study, we retrieve the dataset of Ruiz-Zárate and Arias-González (2004), collected during 1999 and 2000, including richness and distribution of fish and coral community, and the reef zones of the eastern coast of the Yucatán Peninsula, belonging to the north sector of MBRS (Fig. 1). By combination of these data and new information on diet and morphology of the fish community, we analyzed and classified into functional groups the coral reef fish community of the eastern coast of the Yucatán Peninsula. The aims of our study were: *1*) determine the

functional relationships of the fish community based on diet and morphology, and 2) determine the extent of the influence of habitat complexity and the protection of the Sian-Ka'an Biosphere Reserve on the richness of fish and the functional diversity of the ecosystem. The information obtained from our study will provide ecological tools for managers and politician to implement governance measures for the resilience of the MBRS.

Materials and methods

This study considered 11 coral reefs localities of the eastern coast of the Yucatán Peninsula. Reefs in this area show a gradient of anthropogenic pressure from north to south (Arias-González et al., 2008), with the Sian-Ka'an Biosphere Reserve in the central area (Fig. 1). Moreover, reefs show a clear geomorphologic zoning in 5 zones: lagoon, crest, front, slope (corresponding to the spore and grove), and terrace (Arias-González et al., 2008, 2011; Ruiz-Zárate & Arias-González, 2004); however, the crest zone is not considered in fish or coral censuses because is too shallow. For information on sampling design, see Arias-González et al. (2008) and Ruiz-Zárate and Arias-

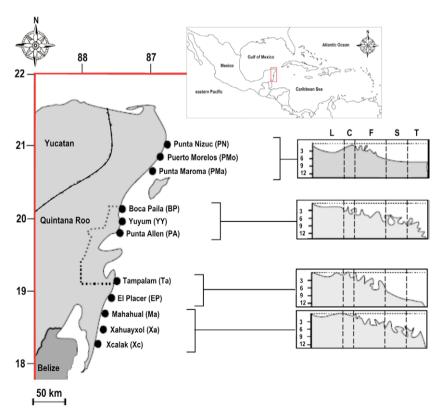


Figure 1. Study sites in the eastern coast of the Yucatán Peninsula, Mexico, belonging to the north sector of the Mesoamerican Barrier Reef. On the right, the 4 types of topographic profiles of the area. Reef division zones: L, reef lagoon; C, reef crest; F, reef front; S, reef slope; T, reef terrace. Numbers indicate depth in meters. The dotted area represents the Sian-Ka'an Biosphere Reserve.

González (2004). Since fish species show trophic and morphological variation throughout their ontogeny, in this study we only considered adults specimens.

To determine the trophic level of each species in a numeric way, we used the troph index available in Fishbase (Froese & Pauly, 2016). Troph index is defined by 5 main categories: i) detritus: debris and carcasses; ii) plants: phytoplankton; iii) zoobenthos: sponges, cnidarians. worms, mollusks, benthic crustaceans, insects, and echinoderms; iv) zooplankton: jellyfish and planktonic crustaceans; and v) nekton: fish (early stages) and cephalopods (Froese et al., 1992). Accordingly, organisms consuming items from the first or/and second groups (detritus and phytoplankton) will present low troph values, while organisms that consume higher energy groups will present highest troph values. Troph index is calculated using qualitative information from the list of items known to occur in the diet of each species and results in a numerical value (Aguilar-Medrano et al., 2015; Costa & Cataudella, 2007).

Two linkage methods of cluster analysis were used to analyze trophic groups: 1) unweighted pair group method with arithmetic mean (UPGMA), using Euclidian distance. UPGMA is an agglomerative hierarchical clustering method that constructs a rooted tree that reflects the structure present in a pairwise similarity matrix (Sokal & Michener, 1958), and 2) Ward's minimum-variance method, which is a clustering algorithm that optimizes based on within-group variance rather than raw distance (Ward, 1963). To choose the best linkage method we used Cophenetic correlation coefficient as measure of the goodness-of-fit of the dendrogram to the original data (Cardini & Elton, 2008; Sokal & Rohlf, 1962). All possible sets obtained from the cluster analyses were examined by multiple analysis of variance (Manova) and analysis of variance (Anova) to test whether the trophic groups were statistically different based on p values.

The mean standard length (SL) of each species was recorded from literature and the relationships between size and trophic values were tested by regression analyses.

On the eastern coast of the Yucatán Peninsula, from 1999 to 2000, 171 species were recorded. From these, 160 were used in this study. The rest were not included due to their morphological variation, which did not allow us to use the same landmark configuration used for the rest of the species. Among the species not included, there are cartilaginous fish (e.g., Ginglymostoma cirratum, Enchelycore nigricans, Dasyatis americana, Aetobatus narinari) and ray finned fishes (e.g., Stygnobrotula latebricola).

A total of 360 images, representing 160 species belonging to 40 fish families were analyzed. The images were collected

from fish collections as SIO (San Diego, CA, USA), LACM (Los Angeles, USA), USNM (Washington, USA) and specialized web resources such as MorphBank (www. morphbank.net) and FishBase (Froese & Pauly, 2016).

Geometric morphometric methods were used to quantify shape and size variation (Bookstein, 1991; Rohlf, 1999; Rohlf & Marcus, 1993; Zelditch et al., 2004) in each trophic group. To capture as much shape variation as possible, 16 landmarks were recorded (Fig. 2). Configurations were optimally superimposed using a Generalized Procrustes analysis (Rohlf, 1999; Rohlf & Slice, 1990) to obtain a matrix of shape coordinates. Relative Warps including both uniform and non-uniform components (Bookstein, 1991; Rohlf, 1993) were calculated from the Procrustes data and used as shape variables for the statistical analyses. Finally, the centroid size (CS) was used as an estimate of size (Bookstein, 1991).

To explore the main axes of shape variation among specimens a principal components analyses (PCA) was performed and to gain a better appreciation of the shape variation associated with the main axes of PCA, thin-plate spline algorithm (Bookstein, 1991) was used to produce transformation grids representing extreme positive and negative deviations along the axis.

First, trophic groups were tested for the presence of overall morphometric differences using one-way Manova and pairwise multiple comparisons using Hotelling's T value with uncorrected significance. Next, the morphological variation within each group was explored by cluster analysis (UPGMA and Ward's method) and PCA to determine the possible morphological sets. Finally, each morphological set was tested by analyses of variance as canonical variance analysis (CVA), Manova, and Anova.

The effect of habitat complexity (coral cover and number of reef zones), and the protection of the reserve (localities inside the Sian-Ka'an Biosphere Reserve vs. outside of the reserve) on fish species richness (number of species) and functional diversity (number of functional groups) per locality were tested by regression analyses.

Geometric morphometric analyses were carrying out in the Tps series (Rohlf, 2016), and univariate and multivariate analyses were computed using PAST 3.11 (Hammer et al., 2001, freely available at http://folk.uio.no/ahammer/past).

Results

The poorest localities in fish richness and coral cover were the northern localities such as Punta Nizuc, Puerto Morelos, and Punta Maroma; also, these localities had only 2 reef zones: lagoon and front. The localities inside the Sian-Ka'an Biosphere Reserve (Boca Paila, Yuyum, Punta

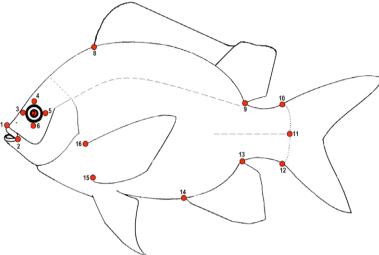


Figure 2. Landmarks used in geometric morphometric analyses: (1-2) snout, (3-7) eye, (8-9) dorsal fin, (10-12) caudal peduncle, (13-14) anal fin, (15-16) insertion of the pectoral fin.

Allen, and Tampalam) had mean values of fish richness and coral cover. Finally, the southern areas (El Placer, Mahahual, Xahuayxol, and Xcalac) had the highest values of fish richness and coral cover (Table 1). From the 40 fish families included in this study, the most diverse families were: Serranidae with 20 species, Haemulidae with 15 species, Labridae with 13 species, and Pomacentridae and Scaridae with 12 species each, while there are 19 families with only 1 species (Table 2).

Table 1
Variables analyzed by locality. NFS: number of fish species; CC: coral cover; NRZ: number of reef zones; S-KBR: protection of the Sian-Ka'an Biosphere Reserve, 1 indicates inside the S-KBR and 0 outside the S-KBR; NFG: number of functional groups.

Localities	NFS	CC	NRZ	S-KBR	NFG
Punta Nizuc	63	21	2	0	12
Puerto Morelos	52	17	2	0	10
Punta Maroma	68	21	2	0	15
Boca Paila	96	45	4	1	13
Yuyum-Xamach	109	66	4	1	16
Punta Allen	93	49	4	1	14
Tampalam	87	52	4	1	12
El Placer	89	56	4	0	15
Mahahual	111	52	4	0	18
Xahuayxol	90	51	4	0	14
Xcalak	93	42	4	0	17

The cophenetic correlation coefficient (CCC) had a better fit for the UPGMA linkage method with Euclidian distance (CCC = 0.856) than Ward's method (CCC = 0.834). UPGMA clusters segregated species in 4 trophic groups (Fig. 3): a, detritivores and mainly herbivores, few including zoobenthos (2-2.2 trophic value); b, mainly zoobenthos, including ectoparasites and zooplankton (2.4-3.3 trophic value); c, zoobenthos, zooplankton, and nekton (3.5-3.8 trophic value); and d, mainly nekton (4.1-4.5 trophic value; Fig. 4). Groups b and c were the most diverse with 47 and 52 species, respectively. Groups a and d were less diverse with 23 and 38 species, respectively. The Manova showed significant differences among the 4 trophic groups (Wilks' lambda = 7.852E-11; F = 5.94E06 ₆₋₃₁₆; p = 0; Pillai trace = 1; $F = 53_{6-318}$; p = 4.46E-45). A small but significant relationship was found between the trophic index and size ($r^2 = 0.07$; std. error = 4.78; t value = 2.05; p = 0.02).

The morphological variation of the community is mainly related to the size of the eyes, the position of the mouth, the size of the dorsal and anal fins, and the deep or elongation of the body (Fig. 5). The first PC summarized 37% of the variation and segregated species with small eyes, posterior dorsal and anal fins, terminal mouth, and short caudal peduncle (e.g., *Diodon hystrix*; PC1+); from species with elongated dorsal and anal fins and superior mouth (e.g., *Bothus lunatus*; PC1-). The second PC summarized 58% of the variation and segregated elongated species, with superior mouth and small eyes (e.g., *Sphyraena barracuda*; PC2+), from deepbodied species, with large eyes and terminal mouth (e.g., *Aluterus schoepfii*; PC2-).

Table 2 Functional reef fish groups of the coast of the Yucatán Peninsula, México.

Functional group	Family	Species	Trophic value
a1	Ostraciidae	Acanthostracion polygonius	2
2	Monacanthidae	Aluterus schoepfii	2
13	Mugilidae	Mugil curema	2
a4	Acanthuridae	Acanthurus bahianus	2
	Acanthuridae	Acanthurus coeruleus	2
	Acanthuridae	Acanthurus chirurgus	2.1
5	Haemulidae	Haemulon melanurum	2.2
	Kyphosidae	Kyphosus sectatrix	2
	Pomacentridae	Stegastes diencaeus	2
	Pomacentridae	Stegastes partitus	2
	Pomacentridae	Microspathodon chrysurus	2.1
	Scaridae	Cryptotomus roseus	2
	Scaridae	Scarus coelestinus	2
	Scaridae	Scarus coeruleus	2
	Scaridae	Scarus iseri	2
	Scaridae	Scarus taeniopterus	2
	Scaridae	Scarus vetula	2
	Scaridae	Sparisoma atomarium	2
	Scaridae	Sparisoma aurofrenatum	2
	Scaridae	Sparisoma chrysopterum	2
	Scaridae	Sparisoma radians	2
	Scaridae	Sparisoma rubripinne	2
	Scaridae	Sparisoma viride	2
1	Monacanthidae	Cantherhines pullus	2.6
	Monacanthidae	Monacanthus ciliatus	2.7
	Monacanthidae	Aluterus scriptus	2.8
	Monacanthidae	Stephanolepis setifer	2.9
	Monacanthidae	Cantherhines macrocerus	3.1
b2	Ostraciidae	Acanthostracion quadricornis	2.7
	Ostraciidae	Lactophrys bicaudalis	3.2
	Ostraciidae	Lactophrys triqueter	3.3
	Ostraciidae	Lactophrys trigonus	3.3
	Tetraodontidae	Canthigaster rostrata	3.3
	Tetraodontidae	Sphoeroides spengleri	3.3
3	Pempheridae	Pempheris schomburgkii	3.1
b4	Balistidae	Melichthys niger	2.4
	Blenniidae	Hypsoblennius invemar	2.8
	Chaenopsidae	Acanthemblemaria aspera	3.1
	Chaenopsidae	Acanthemblemaria maria	3.4

Table 2. Continued.

Functional group	Family	Species	Trophic value
	Chaetodontidae	Prognathodes aculeatus	3.4
	Chaetodontidae	Chaetodon capistratus	3.4
	Cirrhitidae	Amblycirrhitus pinos	3.2
	Gobiidae	Coryphopterus glaucofraenum	2.7
	Gobiidae	Ctenogobius saepepallens	3.4
	Gobiidae	Elacatinus evelynae	3.4
	Grammatidae	Gramma loreto	3.3
	Haemulidae	Haemulon album	3.3
	Haemulidae	Haemulon macrostomum	3.3
	Haemulidae	Haemulon striatum	3.4
	Holocentridae	Holocentrus adscensionis	3.1
	Holocentridae	Myripristis jacobus	3.4
	Labridae	Xyrichtys splendens	3.2
	Labridae	Halichoeres maculipinna	3.3
	Labridae	Thalassoma bifasciatum	3.3
	Labridae	Clepticus parrae	3.4
	Mullidae	Mulloidichthys martinicus	3.2
	Opistognathidae	Opistognathus aurifrons	3
	Pomacanthidae	Pomacanthus paru	2.8
	Pomacanthidae	Holacanthus ciliaris	3
	Pomacanthidae	Holacanthus tricolor	3
	Pomacanthidae	Pomacanthus arcuatus	3.2
	Pomacentridae	Stegastes variabilis	2.5
	Pomacentridae	Stegastes leucostictus	3.1
	Pomacentridae	Stegastes planifrons	3.3
	Pomacentridae	Stegastes fuscus	3.3
	Pomacentridae	Chromis multilineata	3
	Pomacentridae	Chromis scotti	3.4
	Pomacentridae	Chromis insolata	3.4
	Sciaenidae	Equetus lanceolatus	3.4
	Serranidae	Paranthias furcifer	3.2
1	Apogonidae	Apogon maculatus	3.5
2	Balistidae	Canthidermis sufflamen	3.5
3	Diodontidae	Diodon hystrix	3.7
4	Echeneidae	Echeneis naucrates	3.7
5	Fistulariidae	Fistularia tabacaria	3.7
6	Carangidae	Caranx hippos	3.6
	Chaenopsidae	Hemiemblemaria simulus	3.7
	Gobiidae	Elacatinus oceanops	3.5
	Haemulidae	Haemulon vittatum	3.3

Table 2. Continued.

Functional group	Family	Species	Trophic value
	Labridae	Halichoeres radiatus	3.5
	Labridae	Halichoeres pictus	3.5
	Labridae	Halichoeres cyanocephalus	3.6
	Labridae	Halichoeres garnoti	3.7
	Labridae	Halichoeres poeyi	3.7
	Labridae	Xyrichtys martinicensis	3.5
	Malacanthidae	Malacanthus plumieri	3.7
c7	Balistidae	Balistes vetula	3.8
	Chaetodontidae	Chaetodon striatus	3.5
	Chaetodontidae	Chaetodon ocellatus	3.7
	Gerreidae	Gerres cinereus	3.5
	Grammatidae	Gramma melacara	3.5
	Haemulidae	Haemulon bonariense	3.5
	Haemulidae	Haemulon plumierii	3.8
	Haemulidae	Haemulon sciurus	3.5
	Haemulidae	Haemulon chrysargyreum	3.5
	Haemulidae	Haemulon flavolineatum	3.5
	Haemulidae	Haemulon parra	3.5
	Haemulidae	Anisotremus surinamensis	3.6
	Haemulidae	Anisotremus virginicus	3.6
	Haemulidae	Haemulon carbonarium	3.7
	Holocentridae	Holocentrus rufus	3.5
	Holocentridae	Neoniphon marianus	3.6
	Holocentridae	Sargocentron coruscum	3.6
	Holocentridae	Sargocentron vexillarium	3.6
	Labridae	Bodianus pulchellus	3.6
	Labridae	Bodianus rufus	3.7
	Lutjanidae	Lutjanus synagris	3.8
	Mullidae	Pseudupeneus maculatus	3.7
	Sciaenidae	Equetus punctatus	3.5
	Sciaenidae	Pareques acuminatus	3.6
	Sparidae	Calamus bajonado	3.5
	Sparidae	Calamus pennatula	3.7
	Pomacentridae	Abudefduf saxatilis	3.8
	Pomacentridae	Chromis cyanea	3.7
	Sciaenidae	Odontoscion dentex	3.5
	Serranidae	Hypoplectrus gemma	3.8
	Serranidae	Hypoplectrus guttavarius	3.8
	Serranidae	Hypoplectrus puella	3.7
	Serranidae	Serranus tigrinus	3.5

Table 2. Continued.

Functional group	Family	Species	Trophic value
	Serranidae	Epinephelus adscensionis	3.5
	Serranidae	Epinephelus guttatus	3.8
d1	Aulostomidae	Aulostomus maculatus	4.3
	Belonidae	Tylosurus crocodilus crocodilus	4.4
d2	Balistidae	Balistes capriscus	4.1
d3	Bothidae	Bothus lunatus	4.5
d4	Carangidae	Trachinotus falcatus	4
	Carangidae	Trachinotus goodei	4.3
d5	Chaetodontidae	Chaetodon sedentarius	3.9
d6	Diodontidae	Diodon holocanthus	3.9
d7	Carangidae	Caranx crysos	4.1
	Carangidae	Caranx latus	4.2
	Carangidae	Caranx ruber	4.1
	Carangidae	Caranax bartholomaei	4.5
d8	Megalopidae	Megalops atlanticus	4.5
	Synodontidae	Synodus saurus	4.5
d9	Labridae	Lachnolaimus maximus	4.2
	Sparidae	Calamus penna	4.4
d10	Sphyraenidae	Sphyraena barracuda	4.5
d11	Scombridae	Scomberomorus cavalla	4.4
	Scombridae	Scomberomorus maculatus	4.5
	Scombridae	Scomberomorus regalis	4.5
d12	Haemulidae	Haemulon aurolineatum	4.4
	Lutjanidae	Lutjanus apodus	4.3
	Lutjanidae	Lutjanus griseus	4.2
	Lutjanidae	Lutjanus jocu	4.4
	Lutjanidae	Lutjanus mahogoni	4.3
	Lutjanidae	Ocyurus chrysurus	4
	Serranidae	Hypoplectrus indigo	3.9
	Serranidae	Hypoplectrus nigricans	4
	Serranidae	Hypoplectrus unicolor	4
	Serranidae	Cephalopholis cruentata	4.3
	Serranidae	Cephalopholis fulva	4.1
	Serranidae	Epinephelus striatus	4.1
	Serranidae	Mycteroperca bonaci	4.3
	Serranidae	Mycteroperca interstitialis	4.5
	Serranidae	Mycteroperca phenax	4.5
	Serranidae	Mycteroperca tigris	4.5
	Serranidae	Mycteroperca venenosa	4.5
	Serranidae	Rypticus saponaceus	4.1

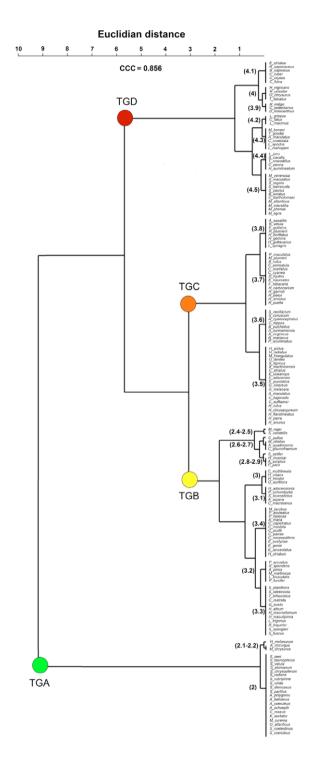


Figure 3. Cluster analysis of the coral reef fish community of the eastern coast of the Yucatán Peninsula, using the trophic index as grouping factor and Euclidian distance. In parenthesis, the trophic values of each cluster. CCC: cophenetic correlation coefficient. Trophic groups: A, green dot; B, yellow dot; C, orange dot; D, red dot.

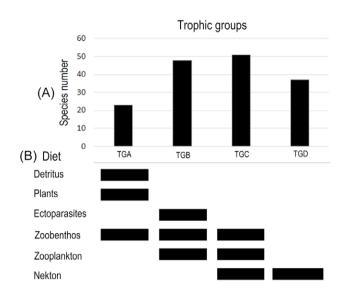


Figure 4. Comparative bar chart of the (A) number of species per trophic group, and (B) diet categories per trophic group. TG: trophic group.

Morphological variation was observed but there was overlap between all trophic groups (Fig. 6). The morphological variation showed how species in trophic group *a* have small eyes and mouth and the eyes are relatively backward from the mouth; the eyes and mouth are larger in group c, which includes highly elongated species (e.g., functional group c3) and deep-bodied species (e.g., functional group c5); trophic group b has deep-bodied species with mean sized eyes and mouth; and finally trophic group d includes the species with the larger mouths and the eyes positioned over the mouth (Fig. 6).

However, Manova results showed significant differences in morphology between the trophic groups (Wilks' lambda = 0.36; F = 4.73_{84-985} ; p = 9.86E-34; Pillai trace = 082; F = 4.47 $_{84-993}$; p = 4.388E-31). By the analyses of shape per trophic group, we found 28 functional groups (Table 2). The trophic group d was the most diverse with 12 functional groups (Wilks' lambda = 3.375E-11; F = 20.24 $_{308-658}$; p = 1.435E-213; Pillai trace = 8.429; F = 8.315 $_{308-781}$; p = 5.048E-126; Fig. 6d), followed by trophic group c with 7 functional groups (Wilks' lambda = 4.429E-05; F = $10.14_{168-379}$; p = 1.217E-76; Pillai trace = 4.301; F = 6.15 $_{168-408}$; p = 8.925E-51; Fig. 6c), trophic group a with 5 functional groups (Wilks' lambda = 9.277E-07; F = 25.79 $_{112-90}$; p = 5.4971E-41; Pillai trace = 3.841; F = 21.59 $_{112-1110}$; p = 2.809E-41; Fig. 6a), and trophic group b with 4 functional groups (Wilks' lambda = 0.0007073; F = 29.82 $_{84-243}$; p = 1.10E-41; Pillai trace = 2.683; F = 25.1 $_{84-249}$; p = 3.85E-85; Fig. 6b).

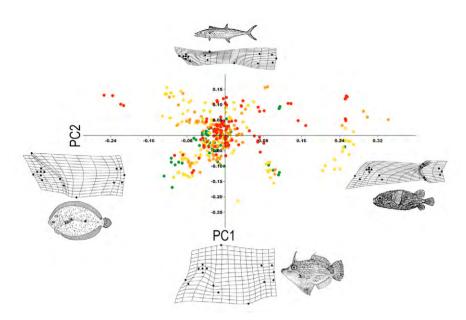


Figure 5. Principal component analysis of the morphometric variation among all species, using PC1 and PC2. Deformation grids indicate extreme shape along PC axes. An illustration was added to exemplify the fish species that each deformation grid represents. Trophic groups: A, green dot; B, yellow dot; C, orange dot; D, red dot.

Functional group c7 is the richest with 34 species, followed by b4 with 33 species, d12 with 18 species, a5 with 17 species and c6 with 10 species, while 13 groups are monospecific (a1, a2, a3, b3, c1, c2, c3, c4, c5, d2, d3, d5, d6, and d10). Seven functional groups were distributed in all localities (a4, a5, b4, c6, c7, d7, d12), 3 were absent only in 1 locality (b1, c1, d10), and 8 were present in only 1 locality (a2 in Xahuayxol; a3 in Mahahual; b3 in Boca Paila; c4 in Tampalam; c5 in Mahahual; d1 in Mahahual; d3 in Xkalak; d5 in Yuyum-Xamach).

A database with 5 variables (coral cover, number of reef zones, protection of the reserve, coded as 1 inside the reserve and 0 outside the reserve, number of species, and number of functional groups per locality; Table 1) was analyzed to determine the effect of habitat complexity and the protection of the Sian-Ka'an Biosphere on the fish and functional diversity. Regression analyses results showed relationships among most variables, except with the variable protection of the Sian-Ka'an Biosphere Reserve. The variables with the strongest relationship were coral cover, number of reef zones, and number of fish species per locality (Table 3).

Discussion

Our results on the coral reef fish community of the eastern coast of the Yucatán Peninsula, belonging to the North Sector of MBRS indicate that variation in habitat complexity was a significant factor explaining fish and functional diversity. Accordingly, a highly complex habitat may support more potential niches and a greater range of trophic groups favoring morphological radiation in its inhabitants and therefore a functionally diverse suite of fauna (Aguilar-Medrano et al., 2012; Arias-González et al., 2011; Fulton, 2007; Klopfer & MacArthur, 1961; Palumbi, 1997).

The Caribbean is one of the 5 world hotspots for marine biodiversity (Rivera-Monroy et al., 2004) and according to our study, a functionally diversified environment. However, resilience depends on many more variables, which allow an ecosystem to resist perturbations and recover (Folke et al., 2004). According to our results, the Caribbean might not be a highly resilient ecosystem because the functional groups show low redundancy (from the 28 functional groups, 22 include 1 to 5 species) and the reef fish community strongly depends on the coral cover. Besides this, the Caribbean had few hard barriers to dispersal, behaving as a single ecosystem, and thus imposing little buffering from biological or physical disturbance (Roff & Mumby, 2012). Thus, in the case of large-scale disturbance, this area might be hardly affected. As has been shown in different studies, the Caribbean Sea is one of the regions where the most serious declines in terms of reef area, health, and fisheries productivity is occurring (Alvarez-Filip et al., 2009; Gardner et al., 2003; Miloslavich & Klein, 2005; Perry et al., 2013).

As found in previous analyses, our results show that habitat complexity is highly related to fish diversity and functional diversity in the Caribbean reefs (Alvarez-Filip et al., 2009; Arias-González et al., 2008; Núñez-Lara & Arias-González, 1998). This is observed clearly when comparing Yuyum-Xamach or Mahahual, the localities with most of the highest values of coral cover, fish diversity, and functional diversity with the presence of specific functional groups for each locality, against Puerto

Morelos, the locality with the lowest values, which only presents the common functional groups, i.e., those present in all localities.

On the other hand, from the 8 localities with 4 reef zones, in 6 of those, the slope is the zone with the highest fish diversity, while in all localities the lagoon is the zone with the lowest fish diversity. Cinner et al. (2016) proposed a new approach to characterize and manage reefs through the classification of reef zones based on environmental

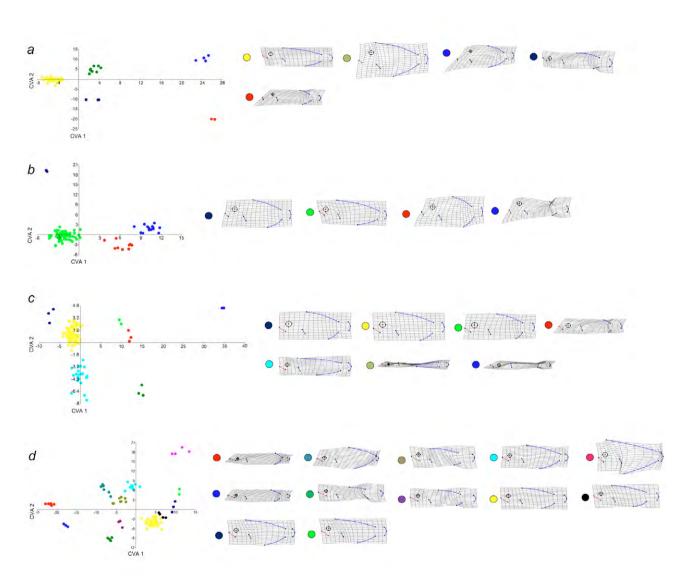


Figure 6. Canonical variables analyses of trophic groups showing the resultant functional groups and their morphological variation. CVAa: a1, red circles; a2, blue circles; a3, navy blue circles; a4, green circles; a5, yellow circles. CVAb: b1, red circles; b2, blue circles; b3, navy blue circles; b4, green circles. CVAc: c1, red circles; c2, blue circles; c3, dark green circles; c4 green circles; c5, navy blue circles; c6, aqua circles; c7 yellow circles. CVAd: d1, red circles; d2, blue circles; d3, pink circles; d4, dark green circles; d5, purple circles; d6, green circles; d7, navy blue circles; d8, aqua circles; d9, brown circles; d10, grey circles; d11, black circles; d12, yellow circles. In the morphological grids (right) marked in black the eye, in red the snout, and in blue the base of the dorsal fin, caudal fin, anal fin, and pectoral fin.

and socioeconomic variables. Accordingly, a bright spot is a zone where the ecosystem is substantially better and a dark spot is a zone where ecosystems are worse. In this sense, according to our results, the slope is a bright spot in all studied localities, without distinguishing between localities inside or outside the reserve. The presence of these bright spots in all studied localities produces a belt of diversity, favoring connectivity. Thus, conservation of these bright spots can ensure the recovery of areas where pressures have produced loss of habitat complexity and diversity.

According to our results, the protection of the Sian-Ka'an Biosphere Reserve does not show a relationship with any of the response variables. Studies on the effect of the marine reserves on fish communities indicate that it is possible to observe results of protection after a few years (Claudet et al., 2006; Halpern & Warner, 2002; Russ et al., 2005) or decades (Claudet et al., 2008; Halpern, 2003; McClanahan & Arthur, 2001; Micheli et al., 2004). Because our data date from 13 and 14 years after the declaration of the marine reserve, time could be a reason for our result and would be useful for testing whether the factors identified here remain important over a longer period.

There is evidence indicating that marine reserves cannot protect reefs from large-scale pollution or global warming (Allison et al., 1998; Hughes et al., 2003; Jameson et al., 2002; Jones et al., 2004). The declaration of marine reserves is gaining importance as a way to protect marine life and avoid anthropogenic impacts, which represents a steppingstone in the protection of these ecosystems. However, despite its popularity, decisions on the design and location have largely been the result of political processes, giving little importance to biological considerations (Agardy, 1994; Halpern, 2003; Jones et al., 1992; McNeill, 1994). Therefore, it is important to reconsider the way we enact and manage marine reserves, including functional approaches that allow us to understand the interactions in the ecosystem, how disturbances affect the structure and function of reefs and permit us to detect key variables to monitor and manage these ecosystems (Bellwood et al., 2004; Hughes et al., 2003).

It is generally accepted that diversity enhances stability in multitrophic systems (Steiner et al., 2005). Our results indicate that the Mexican Caribbean hosts a relatively high diverse conglomerate of fish. However, a community highly dependent on habitat complexity, even with a high functional diversity but low redundancy, as is the case of our study, has the risk of losing functions due to disturbance (Paddack et al., 2009). Functional groups analysis allowed us to have a better idea of how the system works, understanding the trophic and functional

Table 3

Regression analysis on the relation of habitat complexity and protection of the Sian-Ka'an Biosphere Reserve on the coral reef fish community. NFS: number of fish species; NFG: number of functional groups; S-KBR: protection of the Sian-Ka'an Biosphere Reserve; NRZ: number of reef zones; CC: coral cover. Standard error (se). Statistically significant relationships between variables are highlighted in bold.

NFG S-KBR NRZ CC NFS r^2 0.5475 0.1782 0.7896 0.8106 se 1.7489 11.019 3.0150 0.1652 t 3.2998 1.3969 5.8124 6.2065 p 0.0046 0.098 0.0001 0.0001 CC r^2 0.3001 0.2453 0.8541 se 1.905 9.2705 2.2011 t 1.822 1.7105 7.2596 p 0.0487 0.0607 0.0000 NRZ r^2 0.2533 0.2143 se 0.1141 0.5571 t 1.7475 1.5667 p 0.0537 0.0758 S-KBR r^2 0.0211 se 0.0705 t 0.4401 p 0.3351						
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$\begin{array}{c ccccccccccccccccccccccccccccccccccc$		se	1.7489	11.019	3.0150	0.1652
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		p	0.3351			

divisions and by monitoring, detect early changes in the ecosystem.

If we know that coral reefs support over a quarter of the world's small-scale fisheries, generate jobs, provide for many of the 275 million people living close by, and provide critical sea defenses against storms, flooding, and land erosion (Spalding & Brown, 2015), still the pressures over these environments are high. In consequence, it is crucial to develop a better governance by taking efficient actions that regulate the use and management of these ecosystems.

According to our results, we recommend the following measures to protect and manage the ecosystem: *1)* increase efforts in the monitoring of fragile functional groups (locality-specific functional groups) and species-specific functional groups. These groups depend on highly specific niches, thus in the case of disturbance, they might be the first to disappear. Also, increase efforts in the monitoring of functional groups belonging to the trophic group a, which feed mainly on detritus and algae, because their biomass can increase due to coral-algae reefs shift phases

and/or increases in human drainage discharges to the ocean; 2) recognize the slope zone in the Caribbean as a bright zone and accordingly, regulate its use. According to our results, this zone acts as a corridor of diversity for the whole area; 3) compare the functional groups year by year in order to detect early changes in the ecosystem and determine the direction of the change, and 4) educate the fishing community and public in conservation.

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